

# Mangroves in the Gulf of California increase fishery yields

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**Mangroves are disappearing rapidly worldwide despite their well documented biodiversity and the ecosystem services they provide. Failure to link ecological processes and their societal benefits has favored highly destructive aquaculture and tourism developments that threaten mangroves and result in costly "externalities." Specifically, the potentially irreparable damage to fisheries because of mangrove loss has been belittled and is greatly underestimated. Here, we show that, in the Gulf of California, fisheries landings are positively related to the local abundance of mangroves and, in particular, to the productive area in the mangrove–water fringe that is used as nursery and/or feeding grounds by many commercial species. Mangrove-related fish and crab species account for 32% of the small-scale fisheries landings in the region. The annual economic median value of these fisheries is US \$37,500 per hectare of mangrove fringe, falling within the higher end of values previously calculated worldwide for all mangrove services together. The ten-year discounted value of one hectare of fringe is >300 times the official cost set by the Mexican government. The destruction of mangroves has a strong economic impact on local fishing communities and on food production in the region. Our valuation of the services provided by mangroves may prove useful in making appropriate decisions for a more efficient and sustainable use of wetlands.**

discounted values | economic benefits | ecosystem services | small-scale fisheries | *Rhizophora* fringe forest

**M**angrove forests are one of the most biologically important ecosystems in the coastal areas; they contribute to energy flow between land and sea and provide vital ecosystem services, including waste processing, habitat, food production, and recreation (1–3). These coastal forests also strongly influence the structure of neighboring marine communities by increasing the biomass of commercially important fish and invertebrates that spend part of their life cycles in the mangrove environment (4–7). The value of mangrove ecosystem services worldwide has been estimated as an annual global flow of US\$ 1,648 billion (8). Nevertheless, mangroves continue to disappear at an alarming rate because of increased coastal development, tourism, and aquaculture (9–13). This ongoing loss reflects the failure of conservation, fisheries and social scientists, and economists (14, 15) to incorporate ecosystem-based management in public policy. Moreover, evidence linking ecological processes and economic benefits has been scarce (16), and market-oriented strategies for conservation have led to a greatly polarized debate about the paradigm of ecosystem services (17–20). At the same time, economists have used a suite of valuation techniques that have undervalued ecosystem services because of insufficient data and questionable assumptions (21).

The Gulf of California is the northernmost limit for the distribution of mangroves in the Eastern Pacific. On the western coast of the Gulf of California, mangroves are distributed from the Cape region to the center of the Baja California peninsula, mostly in small bays, estuaries, and isolated mangrove pockets (22, 23). On the eastern side of the gulf, mangrove forests are distributed from

Tiburón Island in Sonora, south to Sinaloa and Nayarit, in large coastal lagoons that show extensive mangrove coverage. On the Pacific side of the peninsula, the largest mangrove forests are found inside the coastal lagoons of Magdalena Bay (Fig. 1). Although human density is low in the Gulf of California, there is an increasing pressure to transform mangroves into shrimp farms and tourism developments (24–26). Additionally, modification of water flows through construction of marinas and channels is also an important threat for these ecosystems; for example, coastal areas near La Paz alone lost 23% of the mangrove forests between 1973 and 1981 because of development (27). Today, mangroves are disappearing at a regional rate of 2% annually because of sedimentation, eutrophication, and deforestation (28).

Despite the importance of mangrove forests for food production and economic benefits to local human communities, the potentially irreparable damage to fisheries because of mangrove loss has not been estimated in detail. To test the hypothesis that the amount of mangrove forests has a direct bearing on the production of many commercially important fisheries, in this study we examine the size of fisheries landings in 13 coastal segments of Baja California and the Gulf of California and compare them with the extent of mangrove forests within those same segments. Fisheries data included 9,146 landing records registered between 2001 and 2005 in 25 local offices of the Mexican National Fisheries and Aquaculture Commission (CONAPESCA) in the coastal states of Baja California Sur, Sonora, Sinaloa, and Nayarit (Fig. 1). The majority of the landings recorded in each office came from local fishing grounds that typically include mangrove areas, offshore reefs, or sandy bottoms nearby. Comparable data for the area of mangroves within the same 13 coastal regions were obtained from a wetland database for northern Mexico (29, 30).

## Results

Fisheries landings increased positively with total mangrove area ( $r^2 = 0.70$ ,  $P = 0.0002$ ), but the scaling analysis indicated that a significantly better fit could be obtained if the square root of the mangrove area was used as the predictor of landings ( $r^2 = 0.76$ ,  $P = 0.00004$ ; Fig. 2). The square root of the mangrove area, in turn, was directly related to the length of the mangrove fringe by a simple equivalence relationship that was unrelated to mangrove size or location [fringe = square root of area  $\times$  6.13 ( $\pm 0.45$  SE)]. That is, the fisheries we analyzed only use mangroves as

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**Table 2. Cost estimations of transforming mangrove ecosystems in the Gulf of California**

Time horizon, years	US\$·ha <sup>-1</sup> of fringe mangrove, present discounted values*
4	139,622
6	199,855
10	304,043
30	605,290
50	718,827
100	781,511

\*Per-hectare land value by using a 5% annual discount rate, based exclusively on the long-term contribution of mangroves to regional production of fish and blue crab.

discount rate) of these flows for a hectare of mangrove fringe maintained in productive conditions during a given number of years provide a striking contrast with preceding values set in the region. For instance, over 30 years, the transformation of one hectare of mangrove fringe would cost local economies approximately US\$ 605,290 (Table 2).

## Discussion

Other authors have also found a square-root relationship between mangrove area and fisheries (34), or have stressed the importance of mangrove fringes in the health of coastal ecosystems (35). Our study provides results with implications for public policies designed to regulate the use of coastal resources. First, mangroves in the Gulf of California produce an important amount of food each year. For fish alone, 31.7% of the small-scale fishery landings from 2001 to 2005 comprised species related to mangrove forests. Second, our estimates represent only a lower bound because we considered exclusively the local benefits generated by fish and blue crab without taking into account indirect or existence values (36). Third, although we considered other explanatory environmental variables for fisheries landings, the analysis showed that only the mangrove area is significantly related to the amount of landings produced every year. Finally, in the Mexican government administration time frame (6 years), the fisheries-based long-term value of one hectare of fringe mangrove is 200 times higher than the standard value of US\$ 1,020 ha<sup>-1</sup> established by the Mexican National Forest Commission (37).

Our analysis is based on data for a single time period, and changes in the price of fish and blue crab, or in the harvestable volume of the fisheries, can potentially modify our results. However, considering the price elasticity of demand for fish protein in Mexico (38) and the country's annual population growth of >1%, it seems unlikely that demand for fish will decline significantly in the future. Furthermore, with >60% of the world's fisheries declining (39), it also seems unlikely that there will be dramatic increases in the supply of fish in the near future. Thus, significant decreases in the future value of the fishery services provided by mangroves are improbable.

Our study was done by using a wide-ranging compilation of fisheries landings, data frequently unavailable and very difficult to integrate into an interdisciplinary valuation framework. Additionally, we took advantage of the spatial variation in the amount of mangrove habitat between different fishing regions to develop a more realistic production function of fisheries related with mangroves. These arguments distinguish our economic valuation approach from previous assessments (7, 8, 33), and support the higher value found for a hectare of mangrove just for fisheries alone. Our results highlight the economic benefits of mangrove services to Mexico's economy.

The extreme undervaluation of the benefits generated by mangroves for fisheries versus the projected benefits of coastal

development and aquaculture reveals a management crisis for coastal areas in the Gulf of California. Current agendas pursued by the different economic sectors have been developed independently, resulting in little, if any, compatibility between individual goals, and the undervaluation of ecosystem services in current decision-making processes exacerbates environmental degradation. The precarious state of coastal wetlands in North-western Mexico and in the world in general cannot be ignored, in particular, in an age (40) where food production has important implications for human welfare.

## Materials and Methods

We worked with landing records provided by the National Fisheries Commission (CONAPESCA). Although these records are publicly available by request in any local office in each state's municipalities, data are compiled in CONAPESCA headquarters (Mazatlan, Sinaloa, México), where we obtained a database containing 54,679 records, including monthly crustacean and fish landings. Landings were reported from 2001 to 2005 in 41 offices located along the coasts of the five states surrounding the Gulf of California (Baja California, Baja California Sur, Sonora, Sinaloa, and Nayarit). We only considered data from (Tables S1 and S2): (i) the 25 fisheries offices that have mangrove ecosystems within a 50-km range [small-scale fishing *pangas*, motorized fiberglass skiffs equipped mostly with hand lines and gillnets usually operate within 50 km from their home port (41)]; and (ii) biological groups related to mangroves in any part of their life cycle, such as blue crab, grunts, snappers, snooks, *mojarra*, mullets, and marine catfishes. Data from different offices within the same coastal lagoons were lumped together as part of one single fishing region or corridor, totaling 13 fishing regions (Fig. 1).

We used scaling models to explore how the relationship between mangrove area and fisheries changes as the size of the mangrove habitat increases. Scaling relationships describe how patch-related resource productivity varies with patch size. Ecotonal species that only explore the edge of patches will use their environment as a linear, one-dimensional habitat; as the area of the patch grows, the population numbers will also increase as a function of the square root of the area—a measure of the patch's edge. Conversely, species that use the entire patch area will explore and use their environment fully as a two-dimensional habitat, and their productivity will scale-up linearly with patch area. Thus, the slope of a power function between productivity and patch area plotted on a log-log scale can be used to test hypotheses about the way the resource populations use the environment: if the slope approaches 0.5, then the organisms are mostly using the patch edges; however if the slope approaches unity, the resource species are using the patches in their entirety.

The geographic information used for this study was obtained from an extensive inventory of Mexican wetlands (Table S1), which includes cover data estimated from Landsat Thematic Mapper (TM) satellite images. In particular, we used cover data on (i) mangrove ecosystems (areas dominated by *R. mangle*, *Laguncularia racemosa*, *A. germinans*, and *Conocarpus erectus*), (ii) estuarine systems (deep and shallow waters inside the coastal lagoons), and (iii) submerged aquatic vegetation (deep and shallow waters dominated by bottom-rooted sea-grasses such as *Ruppia maritima*, *Halodule wrightii*, *Syringodium filiformis*, *Zostera marina*, and *Thalassia testudinum*). These three variables can influence the productivity of fish and invertebrates because they represent feeding and nursery habitats.

Because of its large scale, the inventory cited above lacked information on small mangrove wetlands close to four CONAPESCA offices (La Paz, Loreto, Santa Rosalia, and Peñita de Jaltemba). These represent only 0.74% of the mangrove areas considered in this study. It is extremely difficult to work with these isolated mangrove pockets (mean, 14.1 ha ± 2.5 SE) by using satellite images, because the color band of many areas with dense vegetation, predominantly palms or coastal mesquite thickets, could be erroneously confounded with mangrove forests and give an incorrect overestimation of mangrove cover. To solve this problem, we carried out field trips in 2005 and 2006 to survey all mangrove patches distributed in small bays and islands. By using small boats we were able to navigate inside the lagoons and hike in the inland parts of the forests, to take geographic bearings of the extent of mangrove trees. Our field data and observations were used as ground-truthing data points to map the patches as polygons by using Google Earth software. These polygons were saved as KML files (a file format used to display geographic data in an Earth browser such as Google Earth, and that uses a tag-based structure with nested elements and attributes similar to the XML standard), to obtain the total area of the mapped mangrove patches within a 50-km range from each office, by using GE Path software ([www.sgrillo.net](http://www.sgrillo.net)).

To test for alternative hypotheses other than mangrove area explaining variation in landings, we also regressed other potentially explanatory vari-

ables such as the area of sea-grass beds and lagoon estuaries, regional climatic variables such as local precipitation, and fishing effort expressed as the number of boats per fishing region (Table S1). Partial regression coefficients were computed to verify the significance of each variable, and tested by using a two-tailed Student's *t* test.

Finally, to estimate the services provided by mangroves to the local communities, all landings were converted to economic values by using the prices paid locally to the fishermen by fishing cooperatives (the ex-vessel revenue), ignoring all nonfishery benefits of mangroves such as existence and biodiversity values. To remain conservative in our estimates of mangrove value, we also ignored consumer surplus. To get an estimate of the economic worth of

mangroves from the point of view of fisheries, the present discounted values of future annual revenues were accumulated for different time spans, ranging from 4 to 100 years, and by using a 5% discount rate.

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- Blaber SJM (2007) Mangroves and fishes: Issues of diversity, dependence, and dogma. *Bull Mar Sci* 80:457–472.
- Millennium Ecosystem Assessment (2005) *Ecosystems and Human Well-being: Wetlands and Water* (World Resources Institute, Washington, DC).
- Bowman HHM (1917) Ecology and physiology of the red mangrove. *Proc Am Philos Soc* 56:589–672.
- Mumby PJ, et al. (2004) Mangroves enhance the biomass of coral reef fish communities in the Caribbean. *Nature* 427:533–536.
- Dorenbosch M, van Riel MC, Nagelkerken I, van der Velde G (2004) The relationship of reef fish densities to the proximity of mangrove and seagrass nurseries. *Estuar Coast Shelf Sci* 60:37–48.
- Nagelkerken I, et al. (2002) How important are mangroves and seagrass beds for coral-reef fish? The nursery hypothesis tested on an island scale. *Mar Ecol Prog Ser* 244:299–305.
- Rönnebäck P (1999) The ecological basis for economic value of seafood production supported by mangrove ecosystems. *Ecol Econ* 29:235–252.
- Costanza R, et al. (1997) The value of the world's ecosystem services and natural capital. *Nature* 387:253–260.
- Duke NC, et al. (2007) A world without mangroves? *Science* 317:41.
- Barbier EB, Cox M (2003) Does economic development lead to mangrove loss? A cross-country analysis. *Contemp Econ Pol* 21:418–432.
- Alongi DM (2002) Present state and future of the world's mangrove forests. *Environ Conserv* 29:331–349.
- Valiela I, Bowen JL, York JK (2001) Mangrove forests: one of the world's threatened major tropical environments. *Bioscience* 51:807–815.
- Food and Agriculture Organization (2007) *The World's Mangroves 1980-2005* (Food and Agriculture Organization of the United Nations, Rome, Italy).
- Jackson JBC (2007) Economic incentives, social norms, and the crisis of fisheries. *Ecol Res* 22:16–18.
- Hughes TP, Bellwood DR, Folke C, Steneck RS, Wilson J (2005) New paradigms for supporting the resilience of marine ecosystems. *Trends Ecol Evol* 20:380–386.
- Worm B, et al. (2006) Impacts of biodiversity loss on ocean ecosystem services. *Science* 314:787–790.
- McCauley DJ (2006) Selling out on nature. *Nature* 443:27–28.
- Reid WV, et al. (2006) Nature: The many benefits of ecosystem services. *Nature* 443:749.
- Costanza R (2006) Nature: Ecosystems without commodifying them. *Nature* 443:749.
- McCauley DJ (2006) Nature: McCauley replies. *Nature* 443:750.
- Rönnebäck P, Primavera JH (2000) Illuminating the need for ecological knowledge in economic valuation of mangroves under different management regimes—a critique. *Ecol Econ* 35:135–141.
- Contreras-Espinosa F, Warner BG (2004) Ecosystem characteristics and management considerations for coastal wetlands in Mexico. *Hydrobiologia* 511:233–245.
- Whitmore RC, et al. (2006) in *Biodiversity, Ecosystems, and Conservation in Northern Mexico*, eds Cartron JE, Ceballos G, Felger RS (Oxford Univ Press, New York).
- Glenn EP, Nagler PL, Brusca RC, Hinojosa-Huerta O (2006) Coastal wetlands of the northern Gulf of California: inventory and conservation status. *Aquat Conserv* 16:5–28.
- Paez-Osuna F, Guerrero-Galvan SR, Ruiz-Fernandez AC (1999) Discharge of nutrients from shrimp farming to coastal waters of the Gulf of California. *Mar Pollut Bull* 38:585–592.
- Paez-Osuna F, et al. (2003) Shrimp aquaculture development and the environment in the Gulf of California ecoregion. *Mar Pollut Bull* 46:806–815.
- Mendoza R, Amador E, Llinas J, Bustillos J (1984) *Memoria de la primera reunión sobre ciencia y sociedad, presente y futuro de la ensenada de La Paz* (Gobierno del Estado de Baja California Sur, La Paz, México).
- Instituto Nacional de Ecología (2005) *Evaluación preliminar de las tasas de pérdida de la superficie de manglar en México* (Instituto Nacional de Ecología, SEMARNAT, México, D.F.).
- Carrera E, de la Fuente G (2003) *Inventario y clasificación de humedales en México parte I*. (Ducks Unlimited de México A.C., México).
- De la Fuente G, Carrera E (2005) *Cambio de uso del suelo en la zona costera del estado de Sinaloa* (Ducks Unlimited de México A.C., México).
- Manson FJ, Loneragan NR, Harch BD, Skilleter GA, Williams L (2005) A broad-scale analysis of links between coastal fisheries production and mangrove extent: A case-study for northeastern Australia. *Fish Res* 74:69–85.
- Meynecke JO, Lee SY, Duke NC, Warnken J (2007) Relationships between estuarine habitats and coastal fisheries in Queensland, Australia. *Bull Mar Sci* 80:773–793.
- Sathirathai S, Barbier EB (2001) Valuing mangrove conservation in Southern Thailand. *Contemp Econ Pol* 19:109–122.
- Yañez-Arancibia A, et al. (1985) *Fish Community Ecology in Estuaries and Coastal Lagoons: Towards an Ecosystem Integration*, ed Yañez-Arancibia A (UNAM Press, Mexico, DF).
- Barbier EB, et al. (2008) Coastal ecosystem-based management with nonlinear ecological functions and values. *Science* 319:321–323.
- Barbier EB (2000) Valuing the environment as input: Applications to mangrove-fishery linkages. *Ecol Econ* 35:47–61.
- Secretaría de Medio Ambiente y Recursos Naturales (2006) *Acuerdo mediante el cual se expiden los costos de referencia para reforestación o restauración y su mantenimiento para compensación ambiental por cambio de uso de suelo en terrenos forestales y la metodología para su estimación* (Diario Oficial de la Federación DCXXXI-8:11–15).
- Golan A, Perloff JM, Shen EZ (2001) Estimating a demand system with nonnegativity constraints: Mexican meat demand. *Rev Econ Statist* 83:541–550.
- FAO Fisheries and Aquaculture Department (2007) *The State of World Fisheries and Aquaculture 2006* (Food and Agriculture Organization of the United Nations, Rome, Italy).
- Wall G (1998) Implications of global climate change for tourism and recreation in wetland areas. *Clim Change* 40:371–389.
- Sala E, Aburto-Oropeza O, Reza M, Paredes G, Lopez-Lemus LG (2004) Fishing down coastal food webs in the Gulf of California. *Fisheries* 29:19–25.